

Available online at www.sciencedirect.com



Journal of Molecular Catalysis A: Chemical 245 (2006) 235-241



www.elsevier.com/locate/molcata

Contributions of photocatalytic/catalytic activities of TiO_2 and γ -Al₂O₃ in nonthermal plasma on oxidation of acetaldehyde and CO

Taizo Sano*, Nobuaki Negishi, Emiko Sakai, Sadao Matsuzawa

Institute for Environmental Management Technology, National Institute of Advanced Industrial Science and Technology (AIST), AIST Tsukuba West, 16-1 Onogawa, Tsukuba, Ibaraki 305-8569, Japan

> Received 18 August 2005; received in revised form 18 August 2005; accepted 2 October 2005 Available online 10 November 2005

Abstract

The effects of photocatalysis and thermocatalysis in surface discharge type plasma reactors combined with TiO₂- and Al₂O₃-based catalysts were analyzed. The photocatalytic decomposition of acetaldehyde was not observed although the TiO₂ absorbed the UV light ($300 < \lambda < 380$ nm) emitted from N₂ plasma. The intensity of UV emission from plasma decreased with increasing O₂ concentration. In atmospheric air, the photocatalytic decomposition rate by UV light of plasma is lower than 0.2% of the decomposition rate by the plasma itself. The coating of γ -Al₂O₃ on the inside wall of plasma reactor improved the oxidation rate for CO by a factor of 3.5, while TiO₂ improved by only a factor of 1.8, compared with the bare plasma reactor. The high activity of Al₂O₃ seems to be due to the active oxygen species produced by the O₃ decomposition.

© 2005 Elsevier B.V. All rights reserved.

Keywords: Plasma; Photocatalyst; TiO2; Acetaldehyde; CO; Ozone; Surface discharge

1. Introduction

VOCs (volatile organic compounds) are hazardous pollutants emitted from paints, solvents, preservatives, automobile exhaust gas, industrial facilities, etc. [1–5]. Recently, VOCs are also recognized as causative agents of the sick-building syndrome. In addition, emissions of VOCs can contribute to the formation of urban smog and ozone, the stratospheric ozone depletion and the greenhouse effect. Therefore, it is considered that effluent controls at the VOCs emission sources will be severe all over the world. Now, most of the small emission sources of VOCs, such as dry cleaning shop, small printing factory, and construction site using paint, take no effective environmental measures because of cost.

The conventional technique to remove VOCs is activatedcarbon adsorption method, which needs to recycle and dispose the adsorbent used. The recycle of adsorbent requires energy, and the disposal will cause the next environmental problem. The alternative methods are presently being studied widely; catalytic

1381-1169/\$ - see front matter © 2005 Elsevier B.V. All rights reserved. doi:10.1016/j.molcata.2005.10.002

combustion method, nonthermal plasma method, photocatalytic method, ozone oxidation method, etc. [6–22]. Since the catalytic combustion method requires a heat source, it is highly energy-waste process when the VOCs concentration is low. Photocatalytic method using UV-illuminated titanium dioxide (TiO_2) is expected as a safety method with low-energy consumption, since TiO_2 is less toxic powder and the photocatalytic reaction proceeds in air, without any additives, and at ambient pressure [8–10]. Only UV light is required for causing degradation of VOCs. However, the performance of photocatalytic method is too low in the application for VOCs emission sources [11–14]. Alternatively, the photocatalytic method is attractive for the degradation of VOCs with lower concentration such as sick-building syndrome.

The nonthermal plasma (NTP) have been investigated by many researchers for the decomposition of fluorocarbons and VOCs, and found to have excellent performance [6,7,15–22]. Additionally, recently, it is revealed that catalysts in NTP improve the specific energy efficiency and reduce byproduct formation in the VOCs degradation [6,18–22]. In this plasmaenhanced catalyst (PEC) method, many kinds of heterogeneous catalyst, such as BaTiO₃, Al₂O₃, SiO₂, TiO₂, MnO₂ and their derivatives, were evaluated. Among them, TiO₂ often revealed

^{*} Corresponding author. Tel.: +81 29 861 8166; fax: +81 29 861 8258. *E-mail address:* sano-t@aist.go.jp (T. Sano).

the higher enhancement effect on the energy efficiency. Lee et al. reported that the decomposition of benzene by using a discharge plasma-photocatalyst hybrid system using dielectric barrier discharge (DBD) plasma [18]. They described that the photocatalytic reaction by TiO2 increases the benzene conversion and the CO₂ selectivity. Li et al. reported that a combination of dc streamer corona plasma and TiO₂ pellet improved the toluene decomposition [19]. In these reports, the emission of UV light was confirmed but the UV intensities were not analyzed. Additionally, net photocatalytic reaction rates were not evaluated. Ogata et al. reported that the removal rate of fluorocarbons in Ar atmosphere increased when the plasma reactor was filled with TiO₂ pellets as a catalyst [21]. Because significant UV emission and temperature increase were not observed during the plasma operation, they concluded that the catalytic effect seemed to be derived from the direct activation of the TiO₂ surface by the plasma discharge. Thus, photocatalytic effects of TiO_2 in NTP were not well understood. If the photocatalytic effects can improve the specific energy efficiency for the decomposition of VOCs by NTP, the contribution and the theoretical limitation of photocatalytic effects should be discussed. For the purpose, since UV absorption by TiO2 is essential for photocatalytic reactions, the UV intensity emitted from plasma discharge should be analyzed.

In the present paper, we assembled three types of surface discharge plasma reactors equipped with oxide catalysts, and evaluated the contributions of photocatalytic and catalytic effects on the decomposition of VOCs. The intensity of UV emission from plasma was analyzed, and the photocatalytic oxidation rate for acetaldehyde was evaluated. The roles of catalysts, deposited metals, ozone, and heat were discussed using CO oxidation reaction.

2. Experimental

2.1. Preparation of catalysts

Anatase type of TiO₂ (Ishihara Co.: ST-01 and Degssa: P25) and γ -Al₂O₃ (Sumitomo Chemical Co.: TA200I) were used.

The surface areas of ST-01, P25 and TA200I were 300, 50 and 200 m²/g, respectively. Since γ -Al₂O₃ has no photocatalytic activity while it has enough catalytic activities, the effect of γ -Al₂O₃ in plasma was compared with that of TiO₂. Pt-deposited TiO₂ (Pt–TiO₂) and Pt-deposited γ -Al₂O₃ (Pt–Al₂O₃) were also used. Pt–TiO₂ was prepared by photodeposition method with P25 and H₂PtCl₆ as precursors [23]. Pt–Al₂O₃ was prepared by chemical reduction method with H₂PtCl₆ and NaBH₄. The amounts of Pt deposited were 1.0 wt.% against TiO₂ and Al₂O₃. These catalysts were used without further heat-treatment or H₂-reduction.

The catalysts were coated on the reactor walls by the following procedure. A small portion of the catalyst powder was dispersed into distilled water (20 mg-cat/ml-water), and the suspension was painted on the reactor wall while sending a warm current of air to fix the catalyst, and finally the reactor was dried at 110 °C for 12 h. The coated lengths were same as the ground electrode described later.

2.2. Plasma system

The configuration of the basic reactor that generates surface discharge plasma is illustrated in Fig. 1a. A quartz tube (length 180 mm, outer diameter 12 mm, inner diameter 9 mm) was used as the dielectric barrier that played a role of the reactor. A spiral stainless steel wire with a 0.4 mm diameter was set in contact with the inner wall of the barrier tube (15 turns in 100 mm). A copper foil was wrapped on the outside of barrier tube as ground electrode (100 mm length). A high-voltage ac was applied across the electrodes using a 24 kHz neon transformer (S-light Co. Ltd.: model alpha-NEON M-5). The input power (plug-in power) was measured with a digital power meter (Yokogawa Analytical Systems: model WT210). The surface temperatures of the plasma reactor were analyzed by an infrared thermography (Nippon Avionics Co.: TVS-700).

Fig. 1b shows the double tube photocatalytic reactor, where the UV light emitted from plasma was irradiated to the photocatalyst existing out of plasma region. The three windows ($6 \text{ mm} \times 90 \text{ mm}$) were made on the ground electrode, through



Fig. 1. Schematic illustrations of plasma reactors: (a) basic configuration, (b) double tube photocatalytic reactor, (c) plasma enhanced catalytic reactor, and (d) combined reactors.

which the UV light passed. Emission spectra from plasma discharge were measured with an optical spectrograph (Eagle Engineering Co.: LEO/CDI 200–800). TiO₂ powder (P25, 36 mg) was coated on the inner wall of the envelope (60 cm^2) and any catalyst was not coated in the barrier tube. In this case, the effect of photocatalyst would be emphasized since the reactant gas did not contact with plasma. The reactant gas (100 ppm of acetaldehyde diluted in the air) was passed through the space between the inner tube and the envelope at the flow rate of $100 \text{ cm}^3/\text{min}$, and the concentration of acetaldehyde in the effluent was analyzed by a gas chromatograph with FID detector (Shimadzu: GC-14B).

In the analyses of the effects of catalyst in plasma, the catalyst was coated on the inner wall of the barrier tube together with the inner electrode (Fig. 1c). The amount of catalyst was 50 mg. To remove organic impurities on the catalyst surface, plasma was generated at 5 W while clean air was passed through the reactor until the evolution of CO₂ stopped. The reactant gas (800 ppm of CO diluted in Ar/O₂ stream) was passed through the barrier tube at a flow rate of 100 cm³/min. The concentration of CO and CO₂ were analyzed by Fourier transform infrared spectroscopy (FT-IR, Nicolet: Magna650) with a multireflection gas cell (path length 2.4 m). The partial pressures of gases were calculated from the absorbance areas, which were calibrated with standard gases. The concentration of evolved O₃ was analyzed by an ozone analyzer (Ebara Co.: HARE model 620).

Contributions of thermal catalysis in the presence of O_3 were evaluated with the combined reactors shown in Fig. 1d. The reactant gas was same as Fig. 1c. O_3 was formed in the plasma reactor, and was passed through the catalytic reactor, whose temperature was controlled by a thermocouple and an electric furnace. The dimension of catalytic reactor tube was same as the barrier tube.

3. Results and discussion

3.1. Photocatalytic decomposition rate by UV light originated with plasma

The photocatalytic decomposition of acetaldehyde (AcH) by UV radiation from surface discharge plasma was analyzed with the setup (b) in Fig. 1. N₂ gas was passed through the barrier tube, and then high voltage was applied to the electrodes. The input power was controlled to be 5 W. Fig. 2 shows the UV–vis emission spectra observed from the outside of envelope tube. The light emissions with the wavelengths between 290 and 400 nm were observed without TiO₂ coating. This UV-emission range lies within the absorption range of TiO₂. When the envelope with TiO₂ coating was used, the UV light was not detected. This indicates the UV light from plasma was absorbed by TiO₂. Since TiO₂ absorbing UV light generally degrades AcH as Eq. (1) [24], the TiO₂ coated in the envelope can also degrade AcH photocatalytically apart from the problem of its efficiency

$$CH_3CHO + 5/2O_2 \rightarrow 2CO_2 + 2H_2O \tag{1}$$



Fig. 2. UV–vis emission spectra of surface discharge plasma observed from the outside of envelope without catalyst (a) and with TiO_2 (b), and the diffuse reflectance spectrum of TiO_2 . Air was passed through the barrier tube at a flow rate of $100 \text{ cm}^3/\text{min}$. The input power was 8 W.

However, the photocatalytic degradation of AcH by TiO_2 was not observed when 100 ppm of AcH gas was introduced to the reaction space between the barrier tube and the envelope. Fig. 3a shows the relation between the AcH conversion and the input power. AcH conversion was defined as follows:

$$AcH conversion = \left(1 - \frac{AcH concentration at the reactor exit}{\text{supply concentration of AcH}}\right) \times 100$$
(2)

When the input power was 0-3 W, the AcH conversions were zero. By increasing supplied voltage, the AcH conversions were increased. The increases in AcH conversions are due to the plasma and ozone generated at the ground electrode side (Fig. 3b). It seemed that parts of the edges of windows opened on the ground electrode acted as discharge electrodes. However, plasma was mainly generated in the inner electrode side since the concentration of ozone formed on the ground electrode side was much lower than that possibly formed in the inner electrode side. Therefore, it is more suitable to use the ground electrode side for evaluating photocatalytic activity. The AcH conversions with the TiO₂ coating were identical to that without catalyst at the input power range analyzed, and no enhancement effect was observed. This result suggests that the UV intensity of plasma is quite weak to decompose 100 ppm-level of AcH photocatalytically.

When UV light (70 μ W/cm²) of black lamp bulb was irradiated to the TiO₂ from the outside of the envelope, 43% of supplied AcH was decomposed. Since the wavelengths of UV light irradiated were longer than 300 nm, O₃ was not produced. Also the temperature of the photocatalyst was similar to the room temperature. Therefore, the decomposition of AcH was due to only the photocatalytic reaction. The quantum efficiency (incident photon-to-reaction efficiency) was estimated to 28%. In this case, the total UV energy irradiated to TiO₂ was only 4 mW. Considering the result of Fig. 3a, the UV intensities derived from plasma with 3–8 W is estimated to be much smaller than 4 mW.

The intensity of UV light emitted from the surface discharge plasma was measured by GaN type UV sensor (Fuji Xerox Co.:



Fig. 3. Relation between the input power and AcH conversion (a), and O_3 concentration (b), obtained with the double tube photocatalytic reactor. The filled squares (\blacksquare) indicate the values for the experiment without catalyst, and the open circles (\bigcirc) indicate the values with TiO₂ (P25) coated in the envelope.

UV care mate Pro) placed inside of the envelope (Fig. 1b), and was shown in Fig. 4. The diameter of GaN detector was 1.5 mm. The UV intensities were analyzed while controlling the input power and the O₂ concentration in the gas current that was passed through the barrier tube. The total UV intensity was calculated by considering the shapes of the reactor and the window of ground electrode. UV emission was not detected below 2W and the UV intensity increased with the input power above ca. 3 W. The UV intensity was decreased by increasing the O₂ concentration. This suggests that oxygen molecules act as quenching agent for N₂ excitation state. The total UV intensity in the atmospheric condition (O₂ 20%) was only 0.15 mW (ca. 2.5 μ W/cm² on an average) at the input power of 5 W. The photocatalytic reaction rate at the UV irradiation of 0.15 mW is estimated to be $3 \times 10^{-2} \,\mu mol/min$, assuming the quantum efficiency is 100%. On the other hand, the energy efficiencies of plasma for VOCs



Fig. 4. Intensities of UV emission from surface discharge plasma. The total flow rate was $100 \text{ cm}^3/\text{min}$.

decomposition were quite high. The energies to decompose benzene and fluorocarbons are 1×10^7 to 5×10^7 J/mol-VOC in the surface discharge plasma [15,20]. The required energy for AcH decomposition may be less. The potential decomposition rate for AcH by plasma is larger than 20 µmol/min at the input power of 5 W. From these values, the contribution of photocatalysis induced by UV light of the plasma is estimated to be less than 0.2% against the AcH decomposition by the plasma. Since this contribution was obtained by assuming the quantum efficiency of photocatalytic reaction is 100%, the only way to increase the contribution is to increase the UV emission efficiency from plasma dramatically.

3.2. Effects of catalysts existing in plasma

The effects of catalysts, ultrafine TiO₂ and γ -Al₂O₃, in nonthermal plasma on oxidation reaction were investigated. As a model reaction, the oxidation of CO into CO₂ by surface discharge plasma was analyzed. Effects of deposition of Pt on TiO₂ and Al₂O₃ were also analyzed since Pt deposited on TiO₂ enhances the photocatalytic oxidation of CO [24,25]. The reaction atmosphere was Ar/O₂ mixture (50%) containing trace amount of N₂ as illuminant, because plasma in the atmospheric air (N₂/O₂) produces HNO₃, which could deactivate the oxidation sites of TiO₂ [26].

When the surface discharge plasma was generated, the concentration of CO decreased and the concentration of CO_2 increased. Since the decrease in CO concentration was nearly same as the increase in CO_2 concentration, it is considered that CO was oxidized into CO_2 as Eq. (3)

$$\mathrm{CO} + 1/2\mathrm{O}_2 \to \mathrm{CO}_2 \tag{3}$$

The oxidation rate was 29 μ mol/h at the input power of 5 W without catalyst, and the CO conversion was 23%. The CO conversion was improved by all of the catalysts analyzed (Fig. 5a). The enhancement effects by the Al₂O₃ group were larger than



Fig. 5. CO conversions (a) and O_3 concentrations (b) obtained by the plasma enhanced catalytic reactor.

those by the TiO₂ group. Compared with two type of TiO₂, ST-01 (relative surface area = $300 \text{ m}^2/\text{g}$) showed the higher activity than P25 ($50 \text{ m}^2/\text{g}$). This suggests that the surface of catalysts are responsible for the CO oxidation. Also, since the activity of Al₂O₃ ($200 \text{ m}^2/\text{g}$) was higher than that of ST-01, the surface of γ -Al₂O₃ seems to be effective on the CO oxidation rather than that of TiO₂. The deposition of Pt on TiO₂ or Al₂O₃ did not change their activity significantly.

TiO₂ and Al₂O₃ coated in the barrier tube suppressed the concentration of O₃ (Fig. 5b). This is due to (i) the inhibition of O₃ production or (ii) the decomposition of produced O₃. Since O₃ was formed in the plasma without surface, the later reason seems important. The suppression effects of Al₂O₃ group were stronger than those of TiO₂ group. Approximately, the catalyst with stronger O₃-suppression effect showed the higher CO-oxidation activity. The relation between O₃ decomposition and the CO oxidation is discussed later.

In the photocatalytic oxidation of CO, the deposition of Pt particles on TiO₂ increased the CO oxidation rate. It was explained that Pt on TiO₂ acted as the active sites on which CO was chemically adsorbed and was oxidized to CO₂ [25]. In the present paper, the deposition of Pt on TiO₂ did not increase the CO oxidation rate significantly. Therefore, the enhancement of CO oxidation by TiO₂ was not a photocatalytic effect. Also, the increase in the CO oxidation rate by Al₂O₃ is not due to photocatalytic effect because the UV adsorption of Al₂O₃ is weak. The conventional catalytic effect of Al₂O₃ surface or the direct excitation of Al₂O₃ by plasma may be responsible for the enhancement of CO oxidation.

The roles of catalyst and ozone on CO oxidation were further analyzed with the combined reactor shown in Fig. 1d. The thermal reactor composed of a quarts tube and the catalyst was connected at the downstream of the plasma reactor. No catalyst was contained in the plasma reactor. The reaction gas containing CO was passed through the system, and the temperature of the reactor was controlled by a thermoregulator. When the plasma was not generated, Al_2O_3 and TiO_2 did not oxidize CO at the temperatures between 40 and 180 °C (Fig. 6a). Only



Fig. 6. CO conversions by the combined reactors (a) and O_3 concentrations in the outlet gas (b) plotted against the temperature of catalytic reactor. The plasma reactor was operated at 3 W.



Fig. 7. Proposed mechanism of the CO oxidation in the presence of Al_2O_3 and plasma.

Pt–Al₂O₃ oxidized CO at the temperatures above 80 °C. This can be ascribed to the catalytic oxidation activity of Pt with O₂. When the plasma was generated at 3 W and O₃ was supplied to the catalytic reactor, Pt–Al₂O₃ and Al₂O₃ oxidized CO even at the room temperature. The oxidation activity of TiO₂ with O₃ was smaller, and CO was not oxidized without catalyst. These results suggest that CO was oxidized by active species formed on surface of Al₂O₃ by the decomposition of O₃. In the experiment using the setup c in Fig. 1, the surface temperature of plasma reactor became 48–83 °C at the input power of 3–8 W. These temperatures are high enough to take advantage of the catalytic activity of Al₂O₃-based catalyst for CO oxidation.

Fig. 6b shows the O_3 concentration in the effluent from the combined reactor at the input power of 3 W. Without catalyst, the O₃ concentration became ca. 1%. This high concentration is due to the high percentage of O₂ in the reaction atmosphere (Ar/O₂ = 50/50). The O₃ concentration was decreased by the catalytic reactor with Al₂O₃, Pt-Al₂O₃, and TiO₂, and was suppressed effectively at higher temperatures. This indicates that catalytic decomposition of O₃ into O₂ took place. The O₃ decomposition activity was stronger in the order of $Pt-Al_2O_3 > Al_2O_3 > TiO_2$. The catalyst with high O_3 decomposition activity coincided with the catalyst with high CO oxidation activity. The O₃ decomposition on the catalyst was accompanied by formations of active species (O^* and O^*_2) [27]. Therefore, it is considered that the active species formed by the O_3 decomposition oxidized CO into CO_2 as shown in Fig. 7.

The possibility of direct activation of catalyst by plasma was evaluated. The surface temperature of plasma reactor was 48 °C at the input power of 3 W. In the presence of plasma (Fig. 1c), the difference between the CO conversions with Al₂O₃ and without catalyst was 35% at 3 W (Fig. 5a). In the absence of plasma (Fig. 1d), the difference was only 15% at 50 °C in the presence of O₃ (Fig. 6a). Thus, the CO conversion by Al₂O₃ in the plasma was significantly larger than that without plasma, assuming that the temperature of Al₂O₃ in the plasma was similar to the temperature of reactor surface. This suggests that the catalyst existing in plasma has some effects other than the conventional thermal catalysis, such as direct activation of the catalyst by plasma discharge. It was reported that the surface oxygen (or lattice oxygen) of TiO₂ or α -Al₂O₃ is activated by the effect of plasma discharge and is released from the solid phase [21]. If the activity of surface oxygen of γ -Al₂O₃ is also increased by plasma, the oxidation of CO and the decomposition of O₃ are possibly enhanced by the activated surface oxygen. To clarify the behavior of the surface oxygen in plasma, the precise analysis of temperature and isotope techniques are effective.

4. Conclusion

- The contribution of photocatalysis induced by UV light from the surface discharge plasma on the AcH decomposition by the plasma is estimated to be less than 0.2%.
- The coating of γ -Al₂O₃ on the inside wall of plasma reactor improved the oxidation rate dramatically. The high activity of Al₂O₃ is due to the catalytic activity that utilizes O₃ and heat. The direct activation of the surface oxygen by plasma was also suggested.
- To improve energy efficiency of the VOCs decomposition by plasma method, improvements of catalysts that utilize O₃ and lattice oxygen are more important than improvements of photocatalyst activated by weak UV light derived from the plasma.

References

- [1] A.P. Jones, Atmos. Environ. 33 (1999) 4535.
- [2] P. Wargocki, Z. Bako-Biro, G. Clausen, P.O. Fanger, Energy Build. 34 (2002) 775.
- [3] P. Wolkoff, G.D. Nielsen, Atmos. Environ. 35 (2001) 4407.
- [4] P.O. Fanger, Int. J. Refrigerat. 24 (2001) 148.
- [5] E. Racciatti, J. Vecchiet, A. Ceccomancini, F. Ricci, E. Pizzigallo, Sci. Total Environ. 270 (2001) 27.
- [6] S. Futamura, A. Zhang, H. Einaga, H. Kabashima, Catal. Today 72 (2002) 259.
- [7] D. Evans, L.A. Rosocha, G.K. Anderson, J.J. Coogan, M.J. Kushner, J. Appl. Phys. 74 (1993) 5378.
- [8] M.R. Hoffmann, S.T. Martin, W. Choi, D.W. Bahnemann, Chem. Rev. 95 (1995) 69.
- [9] M.D. Driessen, A.L. Goodman, T.M. Miller, G.A. Zaharias, V.H. Grassian, J. Phys. Chem. B 102 (1998) 549.
- [10] K.I. Zamaraev, M.I. Khramov, V.N. Parmon, Catal. Rev. 35 (1994) 617.
- [11] T. Sano, N. Negishi, S. Kutsuna, K. Takeuchi, J. Mol. Catal. A: Chem. 168 (2001) 233.
- [12] H. Einaga, S. Futamura, T. Ibusuki, Phys. Chem. Chem. Phys. 1 (1999) 4903.
- [13] E. Obushi, T. Sakamoto, K. Nakano, F. Shiraishi, Chem. Eng. Sci. 54 (1999) 1525.
- [14] J.L. Falconer, K.A. Magrini-Bair, J. Catal. 179 (1998) 171.
- [15] H.H. Kim, Plasma Process. Polym. 1 (2004) 91.
- [16] J. Arno, J.W. Bevan, M. Moisan, Environ. Sci. Technol. 30 (1996) 2427.
- [17] A. Dono, C. Paradisi, G. Scorrano, Rapid Comm. Mass Spectrometry 11 (1997) 1687.
- [18] B.Y. Lee, S.H. Park, S.C. Lee, M. Kang, S.J. Choung, Catal. Today 93/95 (2004) 769.
- [19] D. Li, D. Yakushiji, S. Kanazawa, T. Ohkubo, Y. Nomoto, J. Electrostat. 55 (2002) 311.
- [20] H.H. Kim, Y.H. Lee, A. Ogata, S. Futamura, Catal. Comm. 4 (2003) 347.
- [21] A. Ogata, H.H. Kim, S. Futamura, S. Kushiyama, K. Mizuno, Appl. Catal. B: Environ. 53 (2004) 175.
- [22] S. Futamura, H. Einaga, H. Kabashima, L.Y. Hwan, Catal. Today 89 (2004) 89.

- [23] T. Sano, N. Negishi, K. Uchino, J. Tanaka, S. Matsuzawa, K. Takeuchi, J. Photochem. Photobiol. A: Chem. 160 (2003) 93.
- [24] A.V. Vorontsova, E.N. Savinova, J. Zhensheng, J. Photochem. Photobiol. A: Chem. 125 (1999) 113.
- [25] H. Einaga, M. Harada, S. Futamura, T. Ibusuki, J. Phys. Chem. B 107 (2003) 9290.
- [26] T. Ibusuki, K. Takeuchi, Atmos. Environ. 20 (1986) 1711.
- [27] W. Li, G.V. Gibbs, S.T. Oyama, J. Am. Chem. Soc. 120 (1998) 9041.